

# Radiological Assessment of Copper Mining Wastes from the Iberian Pyrite Belt

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#### Abstract

The Iberian Pyrite Belt (IPB) is in the southwest Iberian Peninsula. Due to heavy mining and processing of sulfide ores, large amounts of historic mining wastes are dispersed along this area, which are called as "legacy sites". Many of these wastes contain large quantities of metals and metalloids (e.g. As, Cd, Cu) including some potentially Naturally Occurring Radioactive Materials (NORM), which can constitute an important environmental threat and a substantial potential public health. Therefore, a novel radioactive and physicochemical characterization of the most relevant mining wastes was performed, including potential lines of valorization.

**Keywords:** Mining wastes, radioactive characterization, Iberian pyrite belt, radionuclides, radiological hazard index

## Introduction

The Iberian Pyritic Belt (IPB), located in the southwest of the Iberian Peninsula, hosts the largest concentration of polymetallic massive sulfides deposits in the world. This area has been heavily mined for some 5000 years, but was particularly intense during the English period, especially between 1870 and 1930 (Yesares et al. 2015), generating a large amount of mining waste, belonging to the more than 100 mines located in this area, most of them abandoned. These wastes have a high content of metals/metalloids such as As, Cu, Zn, and Pb. Moreover, they are Naturally Occurring Radioactive Materials (NORM) wastes since metals extraction from mined minerals can generate materials with naturally occurring radioactive activity according to the EU regulation. Therefore, its radiological implications have to be evaluated. In addition, the toxic metals, metalloids, and associated contaminants as sulfursuppose a potential environmental risk and problems

for public health, as they are close to towns and sensitive aquatic ecosystems(Álvarez-Valero *et al.* 2009).

Considering the potential environmental and human health problems related to these wastes, the main objective of this work has been to perform a comprehensive radioactive characterization of the different mining wastes located at these mines, as well as the evaluation of their radiological risks.

## Methods

Three mining complexes representative of the mines developed in the IPB were selected for study: 1) Sotiel Coronada (SC), 2) Cueva de la Mora (CM), and 3) Tharsis (TH) (Almodóvar *et al.* 2019). The main processes developed by these mine complexes were extraction of polymetallic ores producing shales rejects (SH), pyrite flotation sludges (FP), roasted pyrite for sulfuric acid production (RP), pyrites leached with sulfuric acid (PY) and slags from minerals smelting (SL) (Tornos *et* 



*Figure 1* Locations of studied mines in the Iberian Pyrite Belt (A), and associated mine waste sample sites at Sotiel Coronada (B), Cueva de la Mora (C) and Tharsis (D). Each colour represents an area with the same type of waste.



*al.* 2000). In Fig. 1 can be seen the locations of mines in relation to the IPB and associated samples.

Regarding the methodology, first the most characteristic waste piles in the area were sampled. One sample (1 kg) was taken per mound, but in case the extent was too large, several samples were taken from that area for better comparison. The mining wastes were collected subsurface with a shovel, eliminating the upper 5 cm most affected by weathering. The samples were stored in hermetically sealed bags. Subsequently, pre-treatment (drying, milling...) and the measurement of physicochemical parameters were carried out. Afterwards, multi-elemental analysis was carried out by ICP-MS/ICP-OES & XRF and radionuclide concentrations measurement by alpha and gamma spectrometry. Radiological risk indices (Ra<sub>eq</sub>, Hex, Hin and Ic; Eke et al. 2024 ; Paschoa & Steinhäusler 2010) were also calculated. Quality control was applied through blanks, replicates, certified reference samples (IAEA-375 & IAEA-327) and intercomparisons.

#### Results

The average concentrations of the major elements in percentage are shown in Fig 2. Each residue has characteristic concentrations for each element. The main elements are Fe, Si and Al. The leached pyrites and flota-

> rectrometry. Radiological ex, Hin and Ic; Eke *et al.* Steinhäusler 2010) were ulity control was applied icates, certified reference & IAEA-327) and inter-

> ntrations of the major age are shown in Fig 2. haracteristic concentrant. The main elements are eached pyrites and flota

tion residues have a high sulfur content, 3-4 orders of magnitude higher than undisturbed soil. The pyritic waste could generate a possible environmental impact from acid mine drainage (AMD) (Moreno-González *et al.* 2022). Also, all wastes have concentrations 3-4 orders of magnitude above typical soil in metals/metalloids such as As, Cu, Zn and Pb, which could lead to high toxicity (Álvarez-Valero *et al.* 2009).

The mean values of natural radionuclide activity concentrations for different wastes are shown in Fig 3 and Fig 4. For the <sup>238</sup>U series, it can be seen that there is secular equilibrium between the radionuclides, and that each waste has distinctive activity concentrations, with slag and leached pyrites having the highest activity concentrations. The observed values for the waste are consistent with those of typical soil (<sup>238</sup>U: 26-82 Bq/kg ; Monty 2001). For the <sup>232</sup>Th decay series and for <sup>40</sup>K the same conclusions can be applied, distinctive values for each residue are those expected for undisturbed soil (<sup>232</sup>Th: 11-84 Bq/kg; <sup>40</sup>K: 25-1650 Bq/kg; Charles 2001).

The radiological risk indices calculated were radium equivalent activity  $(Ra_{eq})$ , the external  $(H_{ex})$  and internal  $(H_{in})$  radiation indices and the activity concentration index for building materials (Ic) (Fig 4). These are defined by the expressions (Eke *et al.* 2024; Paschoa & Steinhäusler 2010):

nignest activity concentration values for the waste are consi of typical soil (<sup>238</sup>U: 26-82 2001). For the <sup>232</sup>Th decay <sup>40</sup>K the same conclusions distinctive values for each re expected for undisturbed so Bq/kg; <sup>40</sup>K: 25-1650 Bq/kg; C

The radiological risk in were radium equivalent act external  $(H_{ex})$  and internal indices and the activity cond for building materials (Ic) (I defined by the expressions (Paschoa & Steinhäusler 2010)

#### Maior elements

*Figure 2* Average concentrations in percentage of the major elements of pyrite flotation (FP), pyrite roasting (RP), smelting slags (SL), leached pyrite (PY) and shale rejects (SH).





*Figure 3*<sup>238</sup>*U series activity concentrations of pyrite flotation (FP), pyrite roasting (RP), smelting slags (SL), leached pyrite (PY) and shale rejects (SH).* 



*Figure 4* <sup>232</sup>*Th series and* <sup>40</sup>*K activity concentrations of pyrite flotation (FP), pyrite roasting (RP), smelting slags (SL), leached pyrite (PY) and shale rejects (SH).* 

)

$$Ra_{eq} (Bqkg^{-1}) = C_U + 1.43C_{Th} + 0.077C_K$$
 (1)

$$H_{ex} = \frac{C_{U}}{370} + \frac{C_{Th}}{259} + \frac{C_{K}}{4810}$$
(2)

$$H_{in} = \frac{C_{U}}{185} + \frac{C_{Th}}{259} + \frac{C_{K}}{4810}$$
(3)

$$I_{\rm C} = -\frac{C_{\rm Ra}}{300} + \frac{C_{\rm Th}}{200} + \frac{C_{\rm K}}{3000}$$
(4)

where  $C_U$ ,  $C_{Th}$ ,  $C_K$  and  $C_{Ra}$  are the activity concentrations of <sup>238</sup>U, <sup>232</sup>Th, <sup>40</sup>K and <sup>226</sup>Ra respectively in Bq/kg. The Ra<sub>eq</sub> values obtained are less than 370 Bq/kg, so they can be

marketed in the USA. The  $H_{ex}$  and  $H_{in}$  values are below the limit value 1, so they do not pose a radiological risk (Eke *et al.* 2024). Likewise, the Ic being less than 1 in all samples, they can be used as building materials without radiological restrictions (Santos *et al.* 2022).

## Conclusions

A multi-elemental analysis of different mining wastes from 3 mines located in the FPI: Sotiel Coronada, Cueva de la Mora and Tharsis has been carried out, with the following conclusions:





tes of radiological risk indices  $Ra_{eq}$  (Bq/kg),  $H_{ex}$ ,  $H_{in}$  and  $I_c$  for pyralling slag (SL), leached pyrite (PY) and slate rejects (SH).

concentrations 3-4 orders above typical soils for ds with high toxicological tes have been radioactively

**Figure 5** Average values of radiological risk indices  $Ra_{eq}$  (Bq/kg),  $H_{ex}$ ,  $H_{in}$  and  $I_c$  for pyrite flotation (PF), pyrite roasting (RP), smelting slag (SL), leached pyrite (PY) and slate rejects (SH).

1. All wastes have concentrations 3-4 orders of magnitude above typical soils for metals/metalloids with high toxicological implications.

with results for activity

- 2. The mining wastes have been radioactively characterised, with results for activity concentrations of natural radionuclides in the range for typical Spanish soils.
- 3. The calculated radiological risk indices demonstrate that the wastes do not pose a radiological risk.
- 4. Finally, these wastes comply with the Spanish regulations required for gamma radiation emitted as construction materials.

#### Acknowledgements

This work has been supported by the project "Diagnosis and proposals for the environmental recovery of areas affected by industrial and mining activities; Implications for the Huelva estuary" (RESTOREHU), Ref.: TED2021-130361B-I00.

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